

1 **Clearance of ingested neonicotinoid pesticide (imidacloprid) in honey bees**
2 **(*Apis mellifera*) and bumble bees (*Bombus terrestris*)**

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5 James E. Cresswell*^{1,2}, François-Xavier L. Robert¹, Hannah Florance¹ and Nicholas
6 Smirnoff¹

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8 ¹Biosciences, College of Life & Environmental Sciences, University of Exeter
9 Geoffrey Pope Building, Stocker Road, Exeter EX4 4QD, United Kingdom

10

11 ²Centre for Pollination Studies, University of Calcutta, 35 Ballygunge Circular Road,
12 Kolkata-700019, India.

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14 *Corresponding author

15 Email: j.e.cresswell@exeter.ac.uk

16 Telephone: +44 1392 763779; FAX: +44 1392 263434

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18 Running title: Pesticide clearance by bees

19

1 **Abstract**

2 BACKGROUND: Bees in agricultural landscapes are exposed to dietary pesticides
3 such as imidacloprid when they feed from treated mass-flowering crops. Concern
4 about the consequent impact on bees makes it important to understand their
5 resilience. In the laboratory, we therefore fed adult worker bees on dosed syrup
6 ($125 \mu\text{g L}^{-1}$ imidacloprid, $98 \mu\text{g kg}^{-1}$) either continuously or as a pulsed exposure and
7 measured their behaviour (feeding and locomotory activity) and whole-body residues.

8
9 RESULTS: On dosed syrup, honey bees maintained much lower bodily levels of
10 imidacloprid than bumble bees ($< 0.2 \text{ ng}$ vs. 2.4 ng imidacloprid per bee). Dietary
11 imidacloprid did not affect the behaviour of honey bees but it reduced feeding and
12 locomotory activity in bumble bees. After the pulsed exposure, bumble bees cleared
13 bodily imidacloprid after 48 hours and recovered behaviourally.

14
15 CONCLUSION: We attribute the differential behavioural resilience of the two species
16 to the observed differential in bodily residues. The ability of bumble bees to recover
17 may be environmentally relevant in wild populations that face transitory exposures
18 from the pulsed blooming of mass-flowering crops.

19
20 **Keywords:** detoxification, ecotoxicology, insecticide, oilseed rape, pulse exposure,
21 recovery

1 **1 INTRODUCTION**

2 Neonicotinoid pesticides (e.g. imidacloprid, clothianidin and thiamethoxam) are
3 widely used for the systemic protection of crops against biting and sucking insect
4 pests¹. Neonicotinoid residues pervade the roots and green tissues of treated
5 plants² but they also appear at trace levels in the nectar and pollen of flowers, which
6 bees consume³. In various laboratory and semi-field trials, dietary neonicotinoids
7 can have harmful sublethal effects⁴ on both honey bees (*Apis mellifera* L.)^{5,6} and
8 bumble bees (*Bombus spp.*)^{7,8}, which has raised concern over the use of
9 neonicotinoids across extensive areas of crops⁹ and the potential threat to valuable
10 pollination services for crops and wild plants¹⁰. Currently, the existence of other
11 detrimental drivers, such habitat degradation and impacts from pathogens¹⁰, and the
12 lack of decisive field trials^{5,11} leaves uncertainty over the relative importance of low-
13 dose dietary exposures from neonicotinoid-treated mass-flowering crops such as
14 oilseed rape (*Brassica napus* L.).

15
16 A further basis for concern is the length of time that bees are exposed to the
17 pesticide in their diet. For example, each field of oilseed rape blooms for several
18 weeks¹² and so some adult bees that forage on a treated crop's flowers could be
19 exposed to dietary pesticide for their entire flightspan, which is about two or three
20 weeks in bumble bees^{13,14} and one week in honey bees^{15,16}. If bees fail to clear
21 ingested pesticide from their bodies, the persistence of even minute daily intakes
22 could eventually build up to harmful levels over successive days. Furthermore,
23 persistence compromises recovery in adult workers whose flight span intersects
24 partially with the bloom of a mass-flowering crop so as extend beyond the crop's
25 flowering. Fields of a mass-flowering crop like oilseed rape typically bloom in

1 synchrony across a landscape ¹² and then flowering subsides, which causes
2 neonicotinoid-exposed bees to shift their foraging to untreated wild flowers ¹⁷ and to
3 thereby experience a ‘pulsed’ exposure. The onset of a pesticide-free diet could
4 enable bees to recover by clearing pesticide from their systems unless the pesticide
5 is persistent. Recovery is fairly rapid in other organisms following a pulsed exposure
6 to imidacloprid. For example, feeding rates in coccinellid beetles (*Serangium*
7 *japonicum*) ¹⁸ and aphids (*Myzus persicae*) ¹⁹ recovered within 24 h, egg production
8 in whitefly (*Bemisia tabaci*) recovered after 48 hours ²⁰ and the behavioural activities
9 of aquatic larvae of *Chironomas* recovered within six days ²¹. What is known about
10 the persistence of ingested neonicotinoids in bees?

11

12 After a single meal of imidacloprid, honey bees clear the pesticide and its
13 metabolites from their body within 24 hours ²² and clearance is achieved principally
14 by metabolic degradation rather than by excretion of the parent compound in bees ²³
15 and other insects ²⁴. The capacity for clearance may explain the recovery of daily
16 food consumption by honey bee colonies after a four-day pulsed exposure to dietary
17 imidacloprid ²⁵. However, the evolution of whole-body toxicant burdens has not
18 previously been studied in bees. We therefore investigated levels of whole-body
19 residues in bees that fed on a neonicotinoid-dosed diet continuously over a period of
20 eight days. Additionally, we studied the rate of clearance and behavioural recovery
21 after a pulsed exposure of several days, which enabled us to address a scientific
22 controversy. Neonicotinoids are neurotoxic and the reversibility of their interactions
23 with their target sites in the insect nervous system is contested ^{26, 27}. If an
24 assimilated pesticide binds persistently to its target receptors, symptoms should
25 persist after dietary exposure ceases. We therefore investigated the bodily levels of

1 ingested imidacloprid in honey bees during a pulsed exposure in conjunction with
2 assays of behavioural recovery. Since equivalent information on bumble bees is
3 lacking, we studied them in parallel.

4

5 **2 METHODS**

6

7 We collected newly eclosed worker honey bees from the brood of a single queen-
8 right colony that was maintained at the University of Exeter. We obtained worker
9 bumble bees (*Bombus terrestris* L.) from a domesticated colony (Koppert B.V.,
10 Berkel en Rodenrijs, Netherlands). Honey bees were placed in cages of 10
11 individuals (0.12 m x 0.10 m x 0.02 m) and bumble bees were placed individually in
12 cages (0.07 m x 0.07 m x 0.035 m). Bees were maintained under semi-controlled
13 conditions: temperature between 23 °C and 27 °C, relative humidity between 21%
14 and 47%; and 12:12 hours of light:darkness. Bees fed *ad libitum* from syrup feeders.
15 For further husbandry details see Cresswell *et al.*²⁸ For acclimatization, newly caged
16 bees fed on undosed syrup for 24 h before the experimental exposures. We
17 estimated the mean fresh mass of honey bees used in our study as 0.14 g (SE =
18 0.01, $n = 6$) and that of bumble bees as 0.19 g (SE = 0.03, $n = 6$).

19

20 Imidacloprid was obtained as a solution in acetonitrile (Dr. Ehrenstorfer, Augsburg,
21 Germany). Acetonitrile was removed by evaporation using a vacuum concentrator
22 (ScanSpeed MaxiVac Beta, Labogene Aps, Lyngø, Denmark) and the imidacloprid
23 was dissolved in the same volume of purified water before being mixed into feeder
24 syrup (Attracker; Koppert B.V., Berkel en Rodenrijs, Netherlands) at a concentration

1 of 125 $\mu\text{g L}^{-1}$, or 98 $\mu\text{g kg}^{-1}$. This dosage was chosen for its known physiological
2 efficacy^{7,28} and not for environmental relevance.

3

4 We subjected bees to one of three treatments over eight days: the control group was
5 fed undosed syrup; the continuous exposure group was fed dosed syrup throughout;
6 the pulsed exposure group was fed dosed syrup for three days and undosed syrup
7 thereafter. Each treatment group comprised three cages of 10 honey bees and 33
8 individually caged bumble bees. Each day, we measured syrup consumption and
9 locomotory activity and collected three individuals for residue analysis (bumble bees
10 were chosen at random and individual honey bees were collected haphazardly, one
11 per cage).

12

13 To quantify locomotory activity, we observed each cage seven times at successive
14 30 minute intervals. On each occasion, each bee was scored as stationary or moving.
15 For bumble bees, we calculated the proportion of the seven observations in which
16 the bee was in motion. For honey bees, we calculated the proportion of bees in
17 motion in each cage at each interval and then calculated the mean of these seven
18 values. While scoring locomotion, the operator was unaware of the cage treatments.

19

20 To quantify the whole-body residue of imidacloprid in each collected bee, it was
21 placed individually in a 2 ml Eppendorf tube (Sarstedt, Leicester, UK) and stored in a
22 freezer at -80°C . To extract imidacloprid, a steel bead (0.4 mm diameter) and 25%
23 methanol (1 ml) was placed in each vial and each was processed in a cooled tissue
24 homogenizer for 5 mins at 25 rpm (TissueLyser, Qiagen, Crawley, UK). The
25 homogenate was centrifuged (17000 g for 5 mins at 4°C) and we collected the

1 supernatant. For each species, the supernatants from the three bees collected on
2 each day were pooled for LC-MS analysis.

3

4 The supernatant was diluted with an equal volume of 25% acetic acid and then
5 subjected to solid phase extraction (SPE). The SPE column (Discovery DSC-18: bed
6 weight = 50 mg; volume = 1 ml; Supelco, Bellefonte, Pennsylvania, USA) was
7 conditioned with methanol (1 ml) and water (1 ml) before 650 μ l of the sample was
8 loaded. The column was washed with water (1 ml) followed by three elutions with
9 methanol (200 μ l). The combined methanol fractions were dried in the vacuum
10 concentrator and stored at -80°C until LC-MS analysis.

11

12 For LC-MS analysis, each sample was re-suspended in a buffer of 25% methanol
13 (400 μ l) and passed through a 0.2 μ m filter and spiked with a reference standard of 1
14 mg l^{-1} of deuterated imidacloprid (Dr. Ehrenstorfer GmbH, Augsburg, Germany).

15 Each was then separated by liquid chromatography (Agilent 1200, Agilent
16 Technologies, Santa Clara, CA, USA) using a reverse phase column (Agilent
17 ZORBAX Rapid Resolution Eclipse Plus C18, Agilent technologies, Santa Clara,
18 USA) interfaced *via* an electrospray ionisation source to a triple quadrupole mass
19 spectrometer (Agilent 6410, Agilent Technologies, Santa Clara, CA, USA) and 10 μ l
20 of sample was injected. Mobile phase A was 0.1% formic acid + 5% acetonitrile and
21 mobile phase B was 0.1% formic acid + 95% acetonitrile. The conditions of elution
22 were: 0 min-0% B, 10 min-100% B, 12 min-100% B, 12.5 min-0% B. The flow rate
23 was 0.3 ml min^{-1} . The source N_2 gas temperature was held at 350°C with a flow of
24 11 l min^{-1} and a nebulizer pressure of 35 psi. The capillary voltage was 4 kV.
25 Fragmentor and collision energy voltages were 40 V and 20 V respectively.

1 Imidacloprid was identified and quantified by selected reaction monitoring (SRM)
 2 using the product ion m/z 209 derived from the precursor ion of m/z 256. The
 3 deuterated imidacloprid was detected using a precursor ion m/z of 260 and a product
 4 ion m/z of 213. The instrument response was linear between 10^{-2} ng and 1 ng
 5 imidacloprid. The amount of imidacloprid in the samples was estimated from the
 6 relative peak areas of unlabelled and deuterated imidacloprid in SRM
 7 chromatograms. We also adjusted for the recovery rate of the extraction method,
 8 which was quantified in a pilot trial in which known concentrations of deuterated and
 9 unlabelled imidacloprid were added to homogenates from undosed bees before
 10 performing the extraction protocol. The recovery rate from honey bee homogenate
 11 was 64% (SE = 1.1%, $n = 3$) and 52% (SE = 2.5%, $n = 3$) from bumble bee
 12 homogenate.

13
 14 The biological half-life of assimilated imidacloprid (T_{half}) in bumble bees was
 15 estimated as $T_{\text{half}} = \ln(2)/k_e$ where k_e is the elimination constant. The elimination
 16 constant²² is given by: $k_e = [\ln(C_1) - \ln(C_2)] / (t_2 - t_1)$ where C_1 and C_2 are the toxicant's
 17 concentration in the bee at times t_1 and t_2 respectively in the post-dose phase of the
 18 pulsed exposure. The level of detection (LOD) was given by: $\text{LOD} = C_0 + 3 \times \text{SE}(C_0)$,
 19 where C_0 is the mean level of imidacloprid detected in the negative control samples
 20 and $\text{SE}(C_0)$ is the standard error of this value²⁹. The level of quantification (LOQ)
 21 was given by: $\text{LOQ} = C_0 + 10 \times \text{SE}(C_0)$. We calculated daily clearance rate (%) as C
 22 $= 100 * [1 - R_D / (I_D + R_{D-1})]$, where R_D denotes the mean whole-body residue on a given
 23 day, I_D denotes the amount of toxicant ingested on that day and R_{D-1} denotes the
 24 whole-body residue level on the previous day. We calculated R_D as the amount of
 25 imidacloprid per dosed bee minus the amount per undosed bee. For the statistical

1 analysis of behavioural effects, we calculated the average response of each
2 experimental replicate across the exposure period (e.g. each cage yielded a single
3 value of the average daily syrup consumption per bee).

4

5 **3 RESULTS**

6

7 In the LC-MS analyses, the LOD was 0.15 ng of imidacloprid per individual for honey
8 bees and 0.10 ng for bumble bees. The LOQ was 0.21 ng for honey bees and 0.16
9 ng for bumble bees.

10

11 Individual honey bees that fed on dosed syrup for eight days ingested a mean of 2.2
12 ng d⁻¹ of imidacloprid (i.e. a total of 17.4 ng) and maintained bodily residues of
13 approximately 0.2 ng (1.4 ng g⁻¹), which were not distinguishable from residues in
14 bees that fed on undosed syrup (paired *t*-test, *t* = 1.34, *df* = 7, *P* > 0.1; Fig 1a). The
15 daily clearance is therefore estimated as $C \approx 100\%$. Mean *per capita* daily rates of
16 feeding (one-tailed *t*-test, *t* = 0.39, *df* = 4, *P* = 0.36) and mean level of activity (one-
17 tailed *t*-test, *t* = 0.29, *df* = 4, *P* = 0.39) did not differ between dosed and undosed
18 bees (Fig 2).

19

20 Based on the mass of syrup consumed and the concentration of imidacloprid, we
21 estimate that individual bumble bees that fed on dosed syrup for eight days ingested
22 a mean of 6.7 ng d⁻¹ of imidacloprid (i.e. a total of 53.8 ng). From the fourth to the
23 eighth days of feeding on dosed syrup, bumble bees maintained bodily residues of
24 approximately 2.4 ng (12.9 ng g⁻¹), which was higher than the level in undosed bees
25 (paired *t*-test, *t* = 10.24, *df* = 4, *P* < 0.001; Fig 1b). The daily clearance rate in

1 bumble bees is therefore estimated as $C = 88\%$ on the first day of ingesting
2 imidacloprid and $C \approx 68\%$ thereafter. Bodily residues were higher in bumble bees
3 than honey bees (paired t -test, $t = 9.77$, $df = 7$, $P < 0.001$). When imidacloprid was
4 removed from their diet, bumble bees eliminated bodily residues after 48 h (Fig 1b)
5 and the biological half life of imidacloprid was $T_{\text{half}} = 10.3$ hours. Dietary imidacloprid
6 reduced mean daily rates of feeding (one-tailed t -test, $t = 3.94$, $df = 53$, $P < 0.001$)
7 and mean daily locomotory activity (one-tailed t -test, $t = 3.05$, $df = 57$, $P = 0.002$) in
8 bumble bees. Bumble bees in the pulsed exposure became more active than the
9 undosed controls the day after the toxicant was removed from their diet (t -test, $t =$
10 4.79 , $df = 20$, $P < 0.001$; Fig 3b) and their feeding rate appeared to recover (Fig 3a).

11

12 **Discussion**

13

14 As previously ²⁸, we found that imidacloprid at a dietary concentration of
15 approximately 100 parts per billion reduced the rates of feeding and locomotory
16 activity in adult worker bumble bees but not honey bees. We attribute this difference
17 to the observed differential in whole-body residues that was evident during the
18 dietary exposure. Specifically, individual honey bees continuously metabolized or
19 otherwise eliminated their daily intake of approximately 2 ng day^{-1} , which is almost
20 half the LD_{50} (48 h oral $LD_{50} \approx 4.5 \text{ ng}$) ⁵. In contrast, bumble bees cleared less than
21 70% of assimilated imidacloprid each day and therefore exhibited a higher level of
22 whole-body imidacloprid. In bumble bees, the correspondence between behavioural
23 recovery and the clearance of bodily residues after dietary intake ceased supports
24 our interpretation that whole-body residues reflect the relative levels of toxicant at the
25 target site, but not necessarily the absolute levels. Specifically, we recognize that

1 the higher whole-body residue levels in bumble bees may have been caused in part
2 by newly ingested syrup in the bee's relatively large honey stomach. Also, our
3 observations do not exclude the possibility that greater target site-sensitivity
4 contributed to the more severe impact on bumble bees but there is currently no
5 evidence to support this, although such variation is known among other insect
6 species^{30,31}.

7
8 We observed that individual bumble bees ingested approximately three times more
9 imidacloprid than individual honey bees over the eight day exposure. Indeed, the
10 greater feeding rate of bumble bees may be the principle cause of their susceptibility
11 rather than a deficiency in detoxification capacity. Once the levels of bodily residues
12 had stabilized, individual bumble bees were capable of clearing about three times
13 more imidacloprid per day than honey bees (i.e. about 7 ng of imidacloprid per day
14 compared to 2 ng per day). Consequently, the relatively high levels of bodily
15 residues in bumble bees were apparently sustained by their relatively high rates of
16 ingesting toxicant. We cannot fully explain why bumble bees consumed so much
17 more syrup than honey bees. The higher food consumption of bumble bees is not
18 attributable solely to body size because their mass was only 40% greater than that of
19 the honey bees in our study. Also, we cannot attribute it to a putative energetic cost
20 of detoxifying imidacloprid because even undosed bumble bees consumed six times
21 more syrup than undosed honey bees. Instead, we speculate that bumble bees
22 metabolized the syrup while maintaining relatively high body temperatures³². We
23 therefore hypothesize that their high energy requirement may predispose bumble
24 bees to impacts from toxicants in nectar.

25

1 We found no evidence that imidacloprid accumulated persistently in either species.
2 In our experiment, adult honey bees cleared imidacloprid at the rate of ingestion,
3 which consistent with the previously reported biological half life of about four hours ²³.
4 Even in bumble bees, we found that bodily residues equilibrated and that the
5 biological half life of imidacloprid was only approximately 10 hours. Our findings
6 have implications for investigators of environmentally relevant impacts of
7 neonicotinoid pesticides on bees. Specifically, it is unrealistic to apply the daily
8 aggregate dose in a single meal ⁶ because it could have a stronger effect than if the
9 same amount of toxicant were ingested gradually over the course of the day, as
10 would happen if the bee foraged normally on flowers with residues in nectar and
11 pollen.

12

13 Even though bumble bees were affected by dietary imidacloprid, they nevertheless
14 cleared the toxicant from their bodies within 48 hours and recovered behaviourally
15 when fed undosed syrup. This finding undermines previous assertions that
16 imidacloprid irreversibly blocks nicotinic acetylcholine receptors (nAChRs) in the
17 central nervous system of insects ^{26, 33}. Similarly, a wide range of physiological
18 evidence contraindicates irreversibility, as follows. The nicotinic acetylcholine
19 receptors (nAChRs) in the insect nervous system are ligand-gated ion channels that
20 are normally activated by a natural neurotransmitter, acetylcholine, but
21 neonicotinoids also act as ligands ³⁴ and so disrupt coordinated nerve activity. The
22 electronegative pharmacophore of neonicotinoids (a nitro or cyano group) interacts
23 with the binding pocket of the pentameric nAChRs through residues in various
24 polypeptide loops that are upstream of loop B ³⁰, within loop C of α subunits ³⁵, and
25 in loop D of β subunits ^{36, 37}. These interactions are not covalent and instead involve

1 hydrogen bonds ³⁶, electrostatic cation- π interactions ^{38, 39} and Van Der Waals
2 interactions ³⁵, which are all relatively weak and therefore reversible. This potential
3 for reversibility is realized in bath-perfusion experiments on isolated neurones ⁴⁰ that
4 show that the depolarisation caused by bathing the cell in imidacloprid is rapidly
5 reversed once the imidacloprid is washed away. Additionally, competitive-
6 displacement experiments show that radio-labelled imidacloprid is displaced from the
7 binding pocket of nAChRs by acetylcholine itself ³¹. The capacity for displacement is
8 confirmed by bath-perfusion electrophysiology, where the depolarising effect of
9 imidacloprid can be reversed by increasing the concentration of the natural
10 neurotransmitter, acetylcholine ³⁴.

11

12 The bees' capacity for recovery lends significance to the pulsed blooming of mass-
13 flowering crops like oilseed rape. Our findings suggest that bees may recover from
14 exposure to dietary imidacloprid once the flowering of the pesticide-treated crop
15 subsides but further research is required to evaluate the role of post-exposure
16 recovery under environmentally realistic conditions.

17

18 We quantified only imidacloprid but we infer that its metabolic derivatives ²³ were
19 cleared by bees over a similar timescale as their parent compound for the following
20 reasons. In honey bees, there was no indication of any dose-dependent effect on
21 behaviour. In bumble bees, behavioural alteration and recovery corresponded to
22 bodily levels of imidacloprid, which contraindicates the proposition that toxic
23 derivatives of imidacloprid have effects separable from those of the parent
24 compound ²³. Because of these considerations, we do not attribute the post-dose
25 increase in the activity level of bumble bees in the pulsed exposure treatment to the

1 delayed production of a stimulatory derivative. Instead, we propose two
2 explanations. First, withdrawal of a cholinergic agonist can increase sensitivity of
3 serotonergic neurones ⁴¹, which are a type that influences flight activity in insects ⁴².
4 Potentially, a similar mechanism of sensitisation may have increased post-exposure
5 locomotory activity in our experimental bees. Alternatively, the heightened activity
6 may be a bee's response to previous intoxication. Social insects such as ants and
7 honey bees exhibit altruistic self-removal ⁴³ whereby diseased individuals leave the
8 colony. We therefore speculate that the heightened activity of bumble bees post-
9 exposure was the result of an intrinsic adaptive response, namely attempted self-
10 removal but this conjecture needs to be substantiated by a demonstration that
11 bumble bees exhibit this behaviour.

12

13 Bumble bees are affected by dietary concentrations of imidacloprid that are far lower
14 than we used in our present experiment ^{7, 28, 44}, which suggests that incomplete
15 clearance of continuously ingested toxicant occurs irrespective of the level of dietary
16 exposure. To us, it is surprising that bumble bees are affected by rates of ingestion
17 in the region of 1 ng of imidacloprid per day ⁷ even though they are capable of
18 clearing 5 ng daily. One possible explanation is that the metabolic degradation of
19 imidacloprid is not fast enough to prevent low levels of toxicant reaching target sites,
20 but further research is required to evaluate this speculation.

21

22 Other insects recover from the effects of imidacloprid after pulsed exposure ^{18,19,45}
23 but we are the first to demonstrate that behavioural recovery from intoxication
24 coincides with bodily clearance of the toxicant. Our observation supports the
25 hypothesis that the interaction between imidacloprid and its target receptors in the

1 bee nervous system is in large part reversible and not persistent as some have
2 asserted ^{26,33}. In separate experiments, we have also observed recovery of brood
3 production by bumble bee (*B. terrestris*) queens after a pulsed exposure to dietary
4 imidacloprid of 14 days at dosages up to 125 µg L⁻¹ (Laycock and Cresswell,
5 unpublished). We therefore anticipate that recovery, in whole or in part, will
6 generalize to various endpoints in this species.

7

8 Our observation that bumble bees can rapidly clear imidacloprid once ingestion
9 ceases lends significance to the pulsed blooming patterns of bee-attractive mass-
10 flowering crops such as oilseed rape. Specifically, the crop's flowering is typically
11 fairly synchronous across a landscape ¹² and the impact of imidacloprid-treated
12 crops on bumble bees may be ameliorated if they recover as the blooming subsides
13 by switching to a diet of untreated wild flowers ¹⁷. Further research is required to
14 establish whether the behavioural recovery that we observed under laboratory
15 conditions means that bumble bees recover their full performance under ecologically
16 relevant conditions.

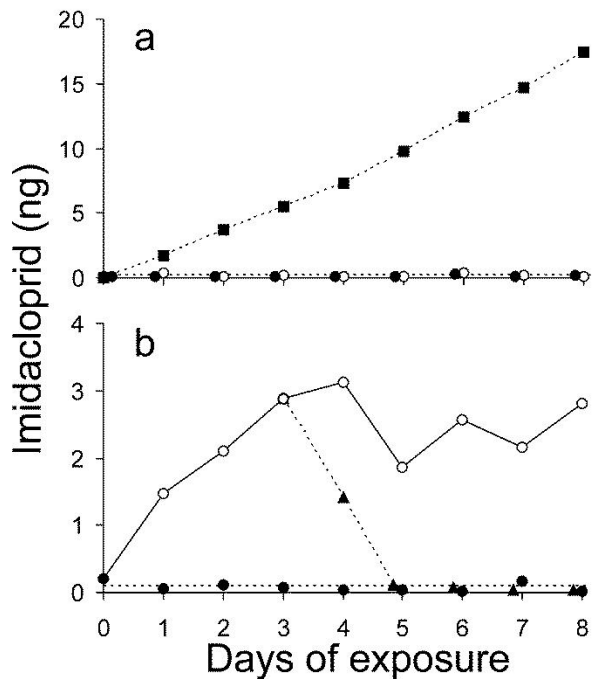
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18 *Acknowledgements.*

19 The work reported here conforms to the regulatory requirements for animal
20 experimentation in the UK and has been approved by the Biosciences Ethics
21 Committee at the University of Exeter.

22

1 FIGURE LEGENDS



2

3 **Figure 1.** Imidacloprid budgets in bees over time in an eight day exposure4 experiment. Panel (a): interpolated square symbols denote the cumulative mass of
5 imidacloprid consumed per honey bee (ng). Other symbols denote whole-body6 residues (ng) in undosed honey bees (filled circles) and dosed honey bees (open
7 circles). The dashed horizontal line indicates the mean whole-body residue in

8 undosed controls calculated across the eight day exposure. Panel (b): whole-body

9 residues (ng) in undosed bumble bees (filled circles) and dosed bumble bees (open
10 circles). Triangles denote the whole-body residues in the pulsed exposure treatment

11 after dosing ceased on day 3. The dashed horizontal line indicates the mean whole-

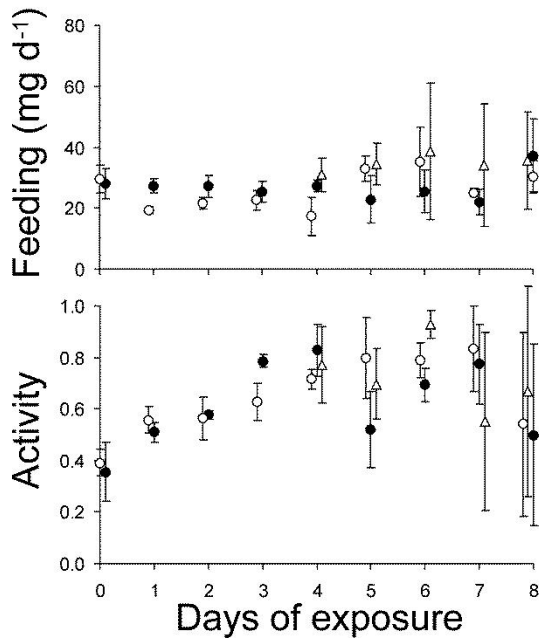
12 body residue in undosed controls calculated across the eight day exposure. Points

13 are interpolated for inspection purposes only and some values are displaced slightly

14 in the x -dimension for ease of inspection. Imidacloprid was assayed in a single

15 pooled homogenate of three individual bees collected from each dose on each day.

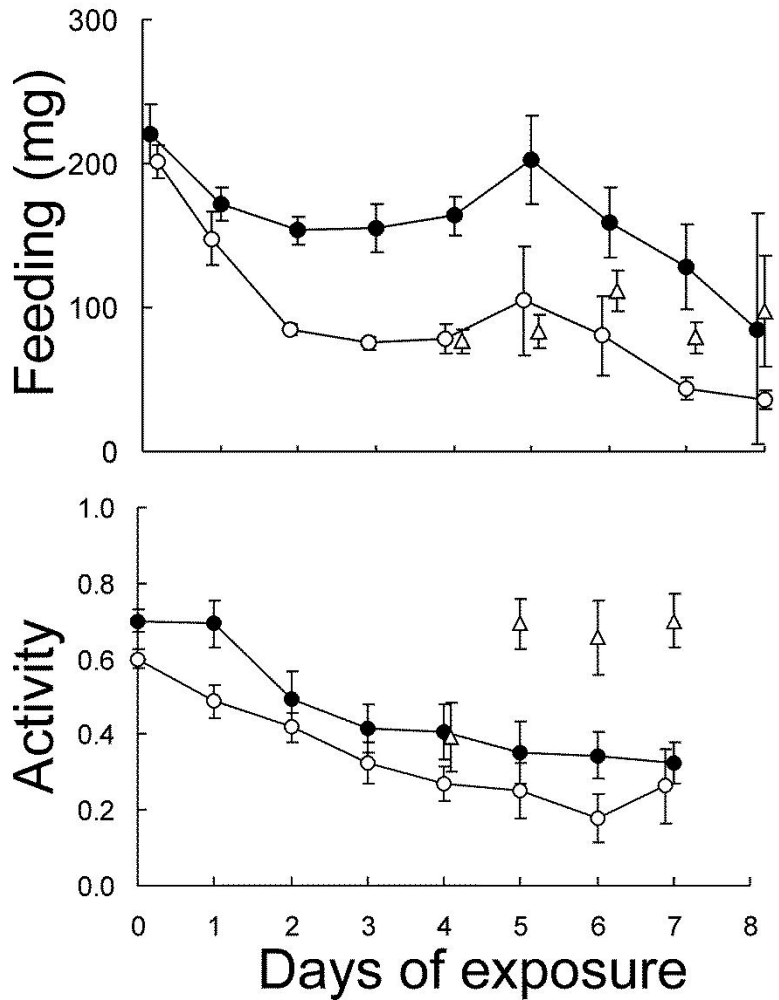
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1

2 **Figure 2.** Behavioural responses (daily feeding rate and locomotory activity) of
3 honey bees over time in an eight day exposure experiment. Panel (a): mass of
4 syrup consumed per bee per day (mg). Panel (b): mean proportion of observations
5 when the individual bee was in motion. Filled circles denote the undosed control
6 treatment; open circles denote the dosed treatment; triangles denote the pulsed
7 exposure treatment after dosing ceased on day 3. Responses on day = 0 are pre-
8 experimental levels. Error bars = 1 SE. Some values are displaced slightly in the x -
9 dimension to reveal their error bars.

10



1
2 **Figure 3.** Behavioural responses (daily feeding rate and locomotory activity) of
3 bumble bees over time in an eight day exposure experiment. Panel (a): mass of
4 syrup consumed per bee per day (mg). Panel (b): mean proportion of observations
5 when the individual bee was in motion. Filled circles denote the undosed control
6 treatment; open circles denote the dosed treatment; triangles denote the pulsed
7 exposure treatment when dosing ceased after day 3. Error bars = 1 SE. Points are
8 interpolated for inspection purposes only and some values are displaced slightly in
9 the x -dimension to reveal their error bars.

10

11

12

1 **REFERENCES**

- 2
- 3 1. Elbert A, Haas M, Springer B, Thielert W and Nauen R, Applied aspects of
4 neonicotinoid uses in crop protection. *Pest Manag Sci* **64**: 1099-1105 (2008).
5
- 6 2. Nauen R and Jeschke P. Basic and applied aspects of neonicotinoid
7 insecticides. In *Green trends in Insect control*, ed. by Lopez O and
8 Fernandez-Bolanos JG. Royal Society of Chemistry: Cambridge, UK, pp. 132-
9 162 (2011).
10
- 11 3. Rortais A, Arnold G, Halm MP and Touffet-Briens F, Modes of honeybees
12 exposure to systemic insecticides: estimated amounts of contaminated pollen
13 and nectar consumed by different categories of bees. *Apidologie* **36**: 71-83
14 (2005).
15
- 16 4. Desneux N, Decourtye A and Delpuech JM, The sublethal effects of
17 pesticides on beneficial arthropods. *Annual Review of Entomology* **52**: 81-106
18 (2007)
19
- 20 5. Cresswell JE, A meta-analysis of experiments testing the effects of a
21 neonicotinoid insecticide (imidacloprid) on honey bees. *Ecotoxicology* **20**:
22 149-157 (2011).
23

- 1 6. Henry M, Beguin M, Requier F, Rollin O, Odoux J-F, Aupinel P, Aptel J,
2 Tchamitchian S and Decourtye A, A common pesticide decreases foraging
3 success and survival in honey bees. *Science* **336**: 348-350 (2012).
4
- 5 7. Laycock I, Lenthall KM, Barratt AT and Cresswell JE, Effects of imidacloprid,
6 a neonicotinoid pesticide, on reproduction in worker bumble bees (*Bombus*
7 *terrestris*). *Ecotoxicology* **21**: 1937-1945 (2012).
8
- 9 8. Whitehorn PR, O'Connor S, Wackers FL and Goulson D, Neonicotinoid
10 pesticide reduces bumble bee colony growth and queen production. *Science*
11 **336**: 351-352 (2012).
12
- 13 9. Environment Agency Austria, *Existing scientific evidence of the effects of*
14 *neonicotinoid pesticides on bees*.
15 [http://www.europarl.europa.eu/committees/en/studiesdownload.html?file=794](http://www.europarl.europa.eu/committees/en/studiesdownload.html?file=79433&languageDocument=EN)
16 [33&languageDocument=EN](http://www.europarl.europa.eu/committees/en/studiesdownload.html?file=79433&languageDocument=EN) [accessed 11 March 2013]
17
- 18 10. Potts SG, Biesmeijer JC, Kremen C, Neumann P, Schweiger O and Kunin WE,
19 Global pollinator declines: trends, impacts and drivers. *Trends Ecol Evol* **25**:
20 345-353 (2010).
21
- 22 11. Blacquière T, Smaghe G, van Gestel CAM and Mommaerts V,
23 Neonicotinoids in bees: a review on concentrations, side-effects and risk
24 assessment. *Ecotoxicology* **21**: 973–992. (2012).
25

- 1 12. McCartney HA and Lacey ME, Wind dispersal of pollen from crops of oilseed
2 rape (*Brassica napus* L.). *Journal of Aerosol Science*; **22**: 467-477 (1991).
3
- 4 13. Brian A, Division of labour and foraging in *Bombus agrorum* Fabricius. *J Anim*
5 *Ecol* **21**: 223-240 (1952).
6
- 7 14. Rodd FH, Plowright RC and Owen RE, Mortality rates of adult bumble bee
8 workers (Hymenoptera: Apidae). *Can J Zool-Rev Can Zool* **58**: 1718-1721
9 (1980).
10
- 11 15. Dukas R, Mortality rates of honey bees in the wild. *Insectes Sociaux* **55**: 252-
12 255 (2008).
13
- 14 16. Rueppell O, Kaftanoglu O and Page R, Honey bee (*Apis mellifera*)
15 workers live longer in small than in large colonies. *Exp Gerontol* **44**: 447-
16 452 (2009).
17
- 18 17. Heinrich B, "Majoring" and "minoring" by foraging bumblebees, *Bombus*
19 *vagans*: an experimental analysis. *Ecology* **60**: 245-255 (1979).
20
- 21 18. He YX, Zhao J, Zheng Y, Zhan Z, Desneux N, Wu KM. Lethal effect of
22 imidacloprid on the coccinellid predator *Serangium japonicum* and sublethal
23 effects on predator voracity and on functional response to the whitefly *Bemisia*
24 *tabaci*. *Ecotoxicology* **21**:1291-1300 (2012)
25

- 1 19. Nauen R, Behaviour modifying effects of low systemic concentrations of
2 imidacloprid on *Myzus persicae* with special reference to an antifeeding
3 response. *Pestic Sci* **44**: 145-153 (1995).
4
- 5 20. He Y, Zhao J, Wyckhuys K and Wu K, Sublethal effects of imidacloprid on
6 *Bemisia tabaci* (Hemiptera: Aleyrodidae) under laboratory conditions. *J Econ*
7 *Entomol* **104**: 833-838 (2011).
8
- 9 21. Azevedo-Pereira H, Lemos M and Soares A, Effects of imidacloprid exposure
10 on *Chironomus riparius* Meigen larvae: linking acetylcholinesterase activity to
11 behaviour. *Ecotox Environ Safe* **74**: 1210-1215 (2011).
12
- 13 22. Suchail S, De Sousa G, Rahmani R and Belzunces LP, *In vivo* distribution
14 and metabolisation of ¹⁴C-imidacloprid in different compartments of *Apis*
15 *mellifera* L. *Pest Manag Sci* **60**: 1056-1062 (2004).
16
- 17 23. Suchail S, Debrauwer L and Belzunces LP, Metabolism of imidacloprid in *Apis*
18 *mellifera*. *Pest Manag Sci* **60**: 291-296 (2004).
19
- 20 24. Wen ZM and Scott JG, Cross-resistance to imidacloprid in strains of German
21 cockroach (*Blattella germanica*) and house fly (*Musca domestica*). *Pestic Sci*
22 **49**: 367-371 (1997).
23
- 24 25. Ramirez-Romero R, Chaufaux J, Pham-Delegue MH, Effects of Cry1Ab
25 protoxin, deltamethrin and imidacloprid on the foraging activity and the

- 1 learning performances of the honeybee *Apis mellifera*, a comparative
2 approach. *Apidologie* **36**: 601-611 (2005).
- 3
- 4 26. Tennekes H, The significance of the Druckrey–Küpfmüller equation for
5 risk assessment - The toxicity of neonicotinoid insecticides to
6 arthropods is reinforced by exposure time. *Toxicology* **276**: 1-4 (2010).
- 7
- 8 27. Maus C and Nauen R, Response to the publication: Tennekes, H.A. (2010):
9 The significance of the Druckrey–Küpfmüller equation for risk assessment—
10 The toxicity of neonicotinoid insecticides to arthropods is reinforced by
11 exposure time. *Toxicology* **280**: 176-177 (2010).
- 12
- 13 28. Cresswell JE, Page C, Uygun M, Holmbergh M, Li Y, Wheeler J, Laycock I,
14 Pook C, de Ibarra N, Smirnoff N and Tyler C, Differential sensitivity of honey
15 bees and bumble bees to a dietary insecticide (imidacloprid). *Zoology* **115**:
16 365-371 (2012).
- 17
- 18 29. Armbruster DA, Tillman MD and Hubbs LM, Limit of detection (LQD)/limit of
19 quantitation (LOQ): comparison of the empirical and the statistical methods
20 exemplified with GC-MS assays of abused drugs. *Clin Chem* 40:1233-1238
21 (1994).
- 22
- 23 30. Liu ZW, Williamson MS, Lansdell SJ, Denholm I, Han ZJ and Millar NS, A
24 nicotinic acetylcholine receptor mutation conferring target-site resistance to

- 1 imidacloprid in *Nilaparvata lugens* (brown planthopper). *P Natl Acad Sci USA*
2 **102**: 8420-8425 (2005).
- 3
- 4 31. Lind RJ, Clough MS, Reynolds SE and Earley FGP, [H-3]imidacloprid labels
5 high- and low-affinity nicotinic acetylcholine receptor-like binding sites in the
6 aphid *Myzus persicae* (Hemiptera : Aphididae). *Pest Biochem Physiol* **62**: 3-
7 14 (1998).
- 8
- 9 32. Heinrich B, Thermoregulation in bumblebees. II. Energetics of warm-up and
10 free flight. *J Comp Physiol* **96**: 155-166 (1975).
- 11
- 12 33. Tennekes H and Sanchez-Bayo F, Time-dependent toxicity of neonicotinoids and
13 other toxicants: implications for a new approach to risk assessment. *J Environ*
14 *Anal Toxicol* **S4**: 001 (2011).
- 15
- 16 34. Deglise P, Grunewald B and Gauthier M, The insecticide imidacloprid is a
17 partial agonist of the nicotinic receptor of honeybee Kenyon cells. *Neurosci*
18 *Lett* **321**: 13-16 (2002).
- 19
- 20 35. Ohno I, Tomizawaa M, Durkin KA, Casida JE and Kagabu S, Bis-
21 neonicotinoid insecticides: Observed and predicted binding interactions with
22 the nicotinic receptor. *Bioorg Med Chem Lett* **19**: 3449-3452 (2009).
- 23
- 24 36. Shimomura M, Yokota M, Ihara M, Akamatsu M, Sattelle DB and Matsuda K,
25 Role in the selectivity of neonicotinoids of insect-specific basic residues in

- 1 loop D of the nicotinic acetylcholine receptor agonist binding site. *Mol*
2 *Pharmacol* **70**: 1255-1263 (2006).
- 3
- 4 37. Bass C, Puinean AM, Andrews M, Cutler P, Daniels M, Elias J, Paul VL,
5 Crossthwaite AJ, Denholm I, Field LM, Foster SP, Lind R, Williamson MS and
6 Slater R, Mutation of a nicotinic acetylcholine receptor beta subunit is
7 associated with resistance to neonicotinoid insecticides in the aphid *Myzus*
8 *persicae*. *BMC Neurosci* **12**: 51.
- 9
- 10 38. Matsuda K, Buckingham SD, Kleier D, Rauh JJ, Grauso M and Sattelle DB,
11 Neonicotinoids: insecticides acting on insect nicotinic acetylcholine receptors.
12 *Trends Pharmacol Sci* **22**: 573-580 (2001).
- 13
- 14 39. Matsuda K, Kanaoka S, Akamatsu M and Sattelle DB, Diverse actions and
15 target-site selectivity of neonicotinoids: structural insights. *Mol Pharmacol* **76**:
16 1-10 (2009).
- 17
- 18 40. Buckingham SD, Lapied B, LeCorronc H, Grolleau F and Sattelle DB,
19 Imidacloprid actions on insect neuronal acetylcholine receptors. *J Exp Biol*
20 **200**: 2685-2692 (1997).
- 21
- 22 41. Rasmussen K and Czachura JF, Nicotine withdrawal leads to increased
23 sensitivity of serotonergic neurons to the 5-HT_{1A} agonist 8-OH-DPAT.
24 *Psychopharmacology* **133**: 343-346 (1997).
- 25

- 1 42. Parker D, Serotonergic modulation of locust motor-neurons. *J Neurophysiol*
2 **73**: 923-932 (1995).
3
- 4 43. Rueppell O, Hayworth MK and Ross NP, Altruistic self-removal of health-
5 compromised honey bee workers from their hive. *J Evol Biol* **23**: 1538-1546.
6
- 7 44. Mommaerts V, Reynders S, Boulet J, Besard L, Sterk G and Smaghe G,
8 Risk assessment for side-effects of neonicotinoids against bumblebees with
9 and without impairing foraging behavior. *Ecotoxicology* **19**: 207-215 (2010).
10
- 11 45. Stoughton SJ, Liber K, Culp J and Cessna A, Acute and chronic toxicity of
12 imidacloprid to the aquatic invertebrates *Chironomus tentans* and *Hyalella*
13 *azteca* under constant- and pulse-exposure conditions. *Arch Environ Con Tox*
14 **54**: 662-673 (2008).